

2 **Inhibition of mussel suspension feeding by surfactants of three classes**3
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Key words: surfactants, filter-feeders, clearance rates, marine mussels, toxicity**Abstract**

Effects of three surfactants on the filtration rates by marine mussels were studied. The xenobiotics tested represented anionic, cationic and non-ionic surfactants (tetradecyltrimethylammonium bromide, a representative of a class of cationic surfactants; sodium dodecyl sulphate, a representative of anionic alkyl sulfates; and Triton X-100, a representative of non-ionic hydroxyethylated alkyl phenols). All three surfactants inhibited the clearance rates. The significance of the results for the ecology of marine ecosystems is discussed.

Abbreviations: CR – clearance rate; EMIS – the electromagnetic induction system; SDS – sodium dodecyl sulphate; TDTMA – tetradecyltrimethylammonium bromide; TX100 – Triton X-100

18 **Introduction**

20 Suspension feeders (filter-feeders) play a significant
21 functional role in aquatic ecosystems. The impor-
22 tant role of filter-feeders (particularly molluscs) is
23 due to their high rates and volumes of water fil-
24 tration (Walz, 1978; Alimov, 1981; Jørgensen
25 et al., 1986; Kryger & Riisgård, 1988; Shulman &
26 Finenko, 1990; Zaika et al., 1990; Dame, 1996). As
27 a result of biological filtration, suspended particles
28 and cells of phytoplankton and microbial plank-
29 ton are removed from the water. This process also
30 accelerates mineralization of organic substances in
31 the filtered matter. Therefore, biological filtration
32 contributes significantly to water purification in
33 aquatic ecosystems.

34 Filter-feeders can accelerate carbon fluxes in
35 ecosystems, because the production of biodeposits
36 (faecal and pseudofaecal pellets) leads to enhanced
37 rates of sedimentation. As a result, bivalves were
38 shown to influence material flux at the sediment-
39 water interface (Smaal et al., 1986; Kautsky &
40 Evans, 1987; Jaramillo et al., 1992; Dame, 1996;
41 Widdows et al., 1998).

Biodeposition rates were estimated as high as 42
60 g m⁻² h⁻¹ at a density of 1400 mussels m⁻² (i.e., 43
50% surface cover) in a mussel (*Mytilus edulis*) 44
bed at Cleethorpes (Humber estuary, England) 45
(Widdows et al., 1998), which is higher than 46
maximum recorded biodeposition rates of 47
25 g m⁻² h⁻¹ for *M. edulis* in the Oosterschelde in 48
the Netherlands (Smaal et al., 1986) and 49
18 g m⁻² h⁻¹ for *M. chilensis* in an estuary in Chile 50
(Jaramillo et al., 1992). Biodeposition rates in 51
some ecosystems were up to 40 times the natural 52
sedimentation rates. Kautsky & Evans (1987) 53
estimated annual biodeposition per g mussel (*M.* 54
edulis, dry weight including shells) as high as 55
1.76 g dry weight, 0.33 g ash-free dry weight, 56
0.13 g carbon, 1.7×10⁻³ g nitrogen and 57
2.6×10⁻⁴ g phosphorus. The annual biodeposition 58
is 11.7 g dry weight per g mussel shell-free dry 59
weight. When average mussel biomass was 60
620 g m⁻² (dry weight including shells) or 91 gm⁻² 61
(dry flesh weight), the annual biodeposition per m² 62
was 1092 g (dry weight), including 80.7 g C, 10.4 g 63
N, 1.6 g P (Kautsky & Evans, 1987). The average 64
composition during the year, expressed as percent 65

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66 of dry weight of biodeposition, was 12.88% C,
67 1.54% N, and 0.19% P. With a total mussel bio-
68 mass of about 10,000 tons in the total 160 km²
69 research area (the northern Baltic proper), the
70 annual contribution from mussels biodeposition
71 would be 1300 tons of carbon, 170 tons of nitro-
72 gen, and 26 tons of phosphorus, which means that
73 the total annual deposition (sedimentation) of C,
74 N, and P is increased by about 10% by mussels as
75 a result of their filtering activity (Kautsky &
76 Evans, 1987).

77 Therefore the measurement of filtration rates is
78 of ecological importance.

79 The marine mussels (*Mytilus edulis*, *M. gallo-*
80 *provincialis* and their hybrids) are important filter-
81 feeders and dominant members of many benthic
82 communities and marine ecosystems. Mussels
83 (*M. edulis*) have been the focus of many studies
84 concerning accumulation of pollutants and their
85 biological effects (Donkin et al., 1989, 1991, 1997).
86 However, few studies have investigated the toxic
87 effects of synthetic surfactants.

88 Generally, surfactants (with very few excep-
89 tions) are not included in the list of priority
90 pollutants (e.g., Scientific Committee for Toxicity
91 and Ecotoxicity of Chemical Substances, Euro-
92 pean Commissions – see Bro-Rasmussen et al.,
93 1994) or are considered of uncertain hazard to the
94 environment. According to Bailey (1996), many
95 surfactants are considered virtually non-toxic for
96 aquatic organisms, provided that the criteria of the
97 Environmental Protection Agency (USA) are
98 valid.

99 Alkyl sulphates, hydroxyethylated alkyl
100 phenols, and quaternary ammonium compounds
101 are important classes of surfactants. Previous re-
102 search has reported inconsistent effects of surfac-
103 tants on certain organisms (Ostroumov, 2000a,
104 2001a) (i.e. both negative and stimulating effects
105 have been recorded). However, at present there is
106 little information on how these surfactants may
107 affect the feeding rate of bivalve molluscs, such as
108 marine mussels.

109 The primary aim of this work was to quantify
110 the effect of surfactants of three classes on the
111 feeding rate of mussels on the algal cell (*Isochrysis*
112 *galbana*). The surfactants studied were sodium
113 dodecyl sulphate (SDS), a representative of alkyl
114 sulfates; Triton X-100 (TX100), a representative
115 of hydroxyethylated alkyl phenols; and tetra-

116 cyltrimethylammonium bromide (TDTMA), a
117 representative of quaternary ammonium com-
118 pounds.

Materials and methods

120 The methods used were similar to those described
121 by Ostroumov (2002a, 2003b). Mussels (*M. edulis*)
122 were collected from a coarse-sand substrate at
123 Exmouth (Devon, England) and used in experi-
124 ments with SDS and TX100. Mussels (natural
125 hybrids *M. edulis*/*M. galloprovincialis*) were
126 collected from the intertidal rocks at Whitsand
127 Bay (Cornwall, England) for use with TDTMA.
128 Mussels were placed in 2-l beakers equipped with
129 magnetic stirrers and kept at 16 °C in a thermo-
130 statically controlled room. Seawater was collected
131 from the Eddystone (~15 km offshore from
132 Plymouth) and filtered through WCN nitrocellu-
133 lose filters with a pore diameter of 0.45 µm
134 (Whatman, Great Britain). A total of 16 animals
135 were studied in each experiment. Eight of them
136 were treated with the xenobiotic, the other eight
137 were controls (no toxicant). The surfactants were
138 added to the experimental beakers 1.5 h before the
139 experiment. The surfactant concentrations shown
140 in the tables and the text are the initial concen-
141 trations of the xenobiotics added to the beakers.

142 In the experiments with SDS and TX100, eight
143 beakers contained eight pairs of mussels with a
144 raw weight of 16–20 g per beaker. An additional
145 beaker containing the 2 l of seawater was used as a
146 reference to confirm that there was no significant
147 change in algal cell concentration in the absence of
148 a mussel and biological filtration of the water.
149 Equal volumes of algal suspension were added
150 simultaneously to the nine beakers.

151 In the experiments with TDTMA, there were 16
152 beakers that each contained one mussel (average
153 wet weight with shell 4.5–5 g). The clearance rate
154 by mussels was determined from the exponential
155 decline in algal cell concentration (*I. galbana*
156 Parke, strain CCAP 927/1). The algal strain was
157 obtained from the NERC Culture Collection
158 of Algae and Protozoa, Dunstaffnage Marine
159 Laboratory, PO Box 3, Oban, Argyll, PA34 4AD,
160 Scotland, UK). Algal cell concentrations were
161 counted with a Coulter Electronics counter
162 (Industrial D model).

Results

164 The clearance rate, or the volume of water cleared
 165 of algal cells per hour, was calculated for each
 166 experimental surfactant concentration and control
 167 condition. The clearance rate at each toxicant
 168 concentration was expressed as a percentage of the
 169 control value (the clearance rate in control mussels
 170 was 100%). The effects of the anionic surfactant
 171 SDS on clearance rate are presented in Table 1.
 172 There was increasing inhibition of clearance rate
 173 with increasing toxicant concentration. The
 174 inhibitory effects of SDS on clearance rate ap-
 175 peared to decline with exposure time during the
 176 course of the 90-min experiment from the moment
 177 labeled as T0 (the beginning of the experiment) to
 178 the moment labeled as T3 (the end of the third
 179 30-min period). These findings are consistent with

Table 1. Inhibition by the anionic surfactant SDS of the mean clearance rates (CR) of mussels (*Mytilus edulis*)

SDS (mg l ⁻¹)	Time period (30 min each)	CR (% of control)
0.5	T0-T1	95.4
	T1-T2	No inhibition measured
	T2-T3	No inhibition measured
1	T0-T1	77.2
	T1-T2	80.8
	T2-T3	88.2
2	T0-T1	55.3
	T1-T2	72.3
	T2-T3	No inhibition measured
4	T0-T1	23.2
	T1-T2	17.9
	T2-T3	30.5
5	T0-T1	4.3
	T1-T2	11.9
	T2-T3	10.3

CR is expressed relative to the control (suspended matter: algae *Isochrysis galbana*) (calculated on the basis of the data of Ostroumov et al., 1997, 1998; Ostroumov, 2001a).

Note. At each of the concentration there were 8 molluscs tested, with 4 experimental and 4 control beakers. There were two molluscs in each of the experimental and control beakers.

the results obtained in another bivalve species (Bressan et al., 1989).

Triton X-I00, a non-ionogenic detergent of the group of hydroxyethylated alkyl phenols, also inhibited the clearance rate by mussels (Table 2).

TDTMA also inhibited the clearance rates of mussels in the experiments with *M. edulis*/*M. galloprovincialis* (Table 3). Clearance rate ceased at 1 mg l⁻¹ and was substantially inhibited at 0.3 mg l⁻¹. The concentrations in the range 0.05–0.3 mg l⁻¹ are similar to those found in the most polluted ecosystems (Review on the Ecological State of Seas, 1992).

Discussion

The levels of surfactants in marine ecosystems often go above maximum permissible concentrations (MAC) reaching levels of >10 MAC and more (Review on the Ecological State of Seas, 1992). In addition, samples of seawater for testing pollutants are usually collected at a distance of >300–500 m from the source of pollution. Therefore, the concentration of pollutants within the area of several hundred meters between the site of sampling and the source of pollution is even higher. This area may include very important coastal ecosystems. Therefore, the results obtained in this study suggest significant deleterious effects caused by environmental levels of surfactants.

Previous studies have shown that the suspension feeding activity of *M. edulis* is also inhibited by other pollutants, including some pesticides (Donkin et al., 1997). Low concentrations of organic pollutants were found to cause a decrease in the feeding rate by *M. galloprovincialis* (Bressan et al., 1989). Furthermore, some commercial detergents (mixtures of several chemicals including surfactants) inhibited water filtering by *M. galloprovincialis* (Ostroumov, 2001a, c).

Similar effects of inhibiting the filtration rate were found when studying effects of TDTMA (0.5 mg l⁻¹) and SDS (0.5 mg l⁻¹) on *Crassostrea gigas* (Ostroumov, 2003b).

Using the electromagnetic induction system (EMIS), it has been shown that several pollutants (copper, cadmium, zinc, lead, tributyltin oxide, chlorine, dispersed crude oil) induced the valve closure response of *M. edulis* (Kramer et al., 1989).

Table 2. Effect of the non-ionic surfactant Triton X-100 (TX100) on the mean clearance rates (CR) of mussels (*Mytilus edulis*) expressed relative to the control (suspended matter: algae *Isochrysis galbana*)

TX100 (mg l ⁻¹)	Time period (30 min each)	CR (+ TX100) l h ⁻¹	CR in control (-TX100) l h ⁻¹	CR (% of control)	Coefficient of inhibition (%)
1	T0-T1	4.04	5.23	77.25	22.75
	T1-T2	4.95	6.13	80.75	19.25
	T2-T3	3.74	4.24	88.21	11.79
2	T0-T1	1.765	4.48	39.42	60.58
	T1-T2	2.77	4.65	59.62	40.38
	T2-T3	2.86	4.72	60.85	39.15
4	T2-T3	0.43	3.02	14.24	85.76
	T3-T4	0.59	1.84	32.06	67.94

Note. At each of the concentration there were 8 molluscs tested, with 4 experimental and 4 control beakers. There were two molluscs in each of the experimental and control beakers.

Table 3. Effect of the cationic surfactant TDTMA on the mean clearance rates (CR) of mussels (*Mytilus edulis*/*M. galloprovincialis*) expressed relative to the control (Suspended matter: algae *Isochrysis galbana*)

TDTMA (mg l ⁻¹)	Time period (50 min each)	CR (+ TDTMA) l h ⁻¹	CR in control (-TDTMA) l h ⁻¹	CR (% of control)
0.05	T0-T1	1.005	1.559	64.47
	T1-T2	1.096	1.290	84.98
	T2-T3	0.936	1.013	92.47
0.1	T0-T1	0.708	1.479	47.84
	T1-T2	0.668	1.383	48.28
	T2-T3	0.455	1.099	41.41
0.3	T0-T1	0.645	1.620	39.82
	T1-T2	0.819	1.640	49.92
	T2-T3	0.350	1.053	33.25
1	T0-T1	0.114	1.168	9.74
	T1-T2	0.100	1.218	8.21
	T2-T3	0.048	0.971	4.89
5	T0-T1	0.051	1.334	3.84
	T1-T2	0.028	1.248	2.20
	T2-T3	0.028	0.871	3.16

Note. At each of the concentration there were 8 molluscs tested, with 8 experimental and 8 control beakers. There was one mollusc in each of the experimental and control beakers.

227 The same toxicants (except crude oil) were detected by the valve closure response of *Dreissena polymorpha* measured by the EMIS (Kramer et al., 1989).

231 Therefore, the effects observed in this study are consistent with the results of other authors. 232 These findings place a particular emphasis on the 233 disturbances to suspension feeders and biofiltration 234 in aquatic systems, a new aspect of the 235 ecological hazard resulting from chemical con-

237 tamination of the environment with surfactants. 238 Water pollution with sublethal concentrations of 239 synthetic surfactants of various classes can inhibit 240 biofiltration in ecosystems, thereby giving rise to 241 additional aspects of ecological hazards (Ostroumov, 2002a, b, 2004). The ecological consequences 242 of biofiltration inhibition may include impairment 243 of water clearance (Ostroumov, 1998), disturbance 244 of biogeochemical fluxes of carbon, increased 245 turbidity and reduced light penetration in the 246

- 247 water column (Ostroumov et al., 1997, 1998) with
 248 adverse effects on phytoplankton and phyto-ben-
 249 thos, as well as other important ecological pro-
 250 cesses in aquatic ecosystems (Ostroumov, 2000b,
 251 2001b, d, e), which provides further evidence in
 252 support of the new approach to prioritization of
 253 anthropogenic effects on biota (Ostroumov,
 254 2003a).
- 255 The ecological significance of the data on
 256 pollutant-induced inhibition of the filtration rate
 257 was discussed in depth in (Ostroumov, 2002c, d).
- 258 Our studies of organisms that are part of
 259 benthic communities ('societies') make us think of
 260 a new interpretation of the poetic words:
- 261 There is society, where none intrudes,
 262 By the deep sea, and music in its roar:
 263 I love not man less, but nature more.
 264 Lord Byron 1788–1824:
 265 *Childe Harold's Pilgrimage* (1812–1818).
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